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Mechanisms, detections, and impacts of species redistributions under climate change

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Abstract

Shifts in species distributions are a common ecological response to climate change, with global temperature rise often hypothesized as the primary driver. However, the direction and rate of distribution shifts are highly variable across species, systems, and studies, complicating efforts to manage and anticipate biodiversity responses to anthropogenic change. In this review, we summarize approaches to documenting species range shifts, discuss why observed range shifts often do not match our expectations, and explore the impacts of species range shifts on nature and society. We use *BioShifts,* a global database of range shift observations, to show that while a majority of documented range shifts are directionally consistent with climate change, large variation remains, with many species not shifting, or shifting in directions opposite from temperature-based expectations. These lagging or expectation-contrary shifts might be explained by additional biotic or abiotic factors driving range shifts, including additional nontemperature climatic drivers, habitat characteristics, and species interactions, all of which are normally unconsidered in range shift documentations. We build the case for three suggested pathways for future research in the field: (i) increasing and connecting observational biological data, (ii) generalizing range shift patterns across systems, (iii) and predicting shifts at management-relevant timescales.

Introduction

Species geographic ranges vary across the tree of life and are dynamic expressions of multiple factors. Abiotic conditions, such a-s temperature and light, are considered to set the bounds of a species' *potential range*: the set of areas where abiotic conditions do not exceed the physiological limits of population sustainability of the species¹. Dispersal limitations, habitat availability, and biotic factors can further limit potential ranges into *realized ranges*: subsets of environmentally suitable areas where a species actually lives^{$2-4$}. While ranges are often illustrated as single units, they are perhaps more usefully conceptualized as populations distributed across space that experience different levels of climatic stress 5 . As temperature

conditions broadly exhibit a large-scale gradient across the globe, species' ranges typically span from a "warm edge" (normally equatorward, downslope in elevation, or shallower in water) to a "cold edge" (normally poleward, upslope, or deeper in water), where conditions experienced by populations can vastly differ (Box 1). Although species might not be directly limited by temperature at either edge and might not be at equilibrium with their environment, populations in these regions will experience different levels of temperature stress and can have different or contrasting responses to temperature change $\mathrm{^{6,7}}.$

Anthropogenic climate change is causing global temperatures to warm on average, sending temperature isolines, or isotherms, shifting across Earth's surface, and altering conditions that organisms experience 8 . At warm edges of species ranges, where individuals are already living near their upper range-wide temperatures, climate warming is expected to cause populations to decrease in growth, reproduction, or survival, diminishing occupancy and contracting the range edge towards the center (i.e. local extinction at the trailing edge). By contrast, at cold edges, where individuals are living near thermal minima, climate warming can relax constraints on organisms (e.g., becoming warm enough for early-life development \degree), allowing for increased survival and persistence of populations and facilitating expansion of the range edge away from its center^{7,10} (Box 1). Together, these changes lead to the expectation that species will shift their distributions towards previously cooler regions as global temperatures warm, leading to range shifts towards higher absolute latitudes, higher elevations in mountains, and deeper waters in $oceans^{11,12}$.

Range shifts are now recognized as a common climate-driven phenomenon with varying impacts on ecosystems, economies, and people $13,14$, yet there is immense variation in the rate and directions of range shifts observed globally^{12,15-19}. Although range-shifting is on one hand a mechanism by which species (and biodiversity) can persist through climate change, an increasingly unstable distribution of life challenges how we interact with, plan to manage, and conserve natural systems. Anticipating range shifts in species of particular ecological, economic, or health concern will be key to successfully adapting to climate change. Moreover, species redistributions and the human responses to them have been linked to a range of social equity imbalances^{20,21}, underscoring understanding of species shifts as an important goal for a sustainable future. Better mechanistic understanding of processes driving and limiting range shifts will increase our ability to respond and learn from them.

In this review, we draw together knowledge of species redistribution processes, summarize the history and methods by which we have detected species' range shifts, and review mechanisms thought to underlie variation in observed rates and directions, from ecological and evolutionary factors to detection methods. We use BioShifts²², a global database including over 26,000 empirical estimates of latitudinal and elevational range shifts compiled from published

literature, to demonstrate general trends, knowledge gaps, and observation biases in range shift documentations. We briefly summarize the impacts of range shifts on natural and human systems as an update to a previous review¹³. We end with a case for areas for improvement in range shift detections and three exciting avenues for future research.

1. Isotherm: isolines connecting areas of equal temperature conditions across space, either at a given instance or as a summary of conditions over time (e.g., annual mean temperature).

2. Range edge: peripheries of a species' geographic distribution. Range edges (also called range margins, or limits) move across space in response to environmental change, and this process is termed "range shifting." Range edges can be difficult to define in space (see Box 2) and may not be set by current climate conditions, but denote the distributional limit for the species.

3. Cold Edge: the edge of a species distribution range located at the colder end of the temperature gradient, often the poleward or upslope edge.

4. Warm Edge: the edge of a species distribution range located at the warmer end of the temperature gradient, often the equatorward or downslope edge.

5. Leading Edge: of a shifting range, the edge experiencing expansions, thus "leading" a range shift. Under climate change, this is normally expected to be the cold edge of the species range, where abiotic constraints will ease with warming.

6. Trailing Edge: of a shifting range, the edge experiencing contractions, thus "trailing" a

range shift. Under climate change, this usually refers to the warm edge of the species range, where warming makes regions of the past range increasingly unsuitable for the species

7. Range Expansion: dispersal and establishment of populations in areas where individuals were not consistently present. Climate change is expected to cause expansions when areas beyond the cold limit warm above the suitable threshold for the species. Also commonly termed "range extension".

8. Range Contraction: population declines or extirpations (i.e., local extinction processes) at the edge of a species range, causing the edge to contract towards the center. In climate change scenarios, contractions are expected to occur, for example, when environments at the warm edge of a species range warm past the upper temperature threshold of the species such that populations are no longer viable.

Section 1 Observations of Climate-Driven Range Shifts

Connections between climate and species distributions have long been documented in the scientific literature²³. Throughout the early-to-mid 1900s, changes in biogeographic ranges were observed in association with local periods of warming and cooling, extreme weather events, and land-use changes (many of these early studies reviewed in 24). These early observations of range changes shed light onto when and where specific factors limit species ranges²⁵. Decades later, the onset of increasingly rapid and consistent global warming (for example, at increased rates beginning in the early 1980s) provided a large-scale– if pseudo-replicated across earth – experiment in which to test the impacts of climate conditions on species redistributions in natural environments 26 .

The first contemporary observations of climate-driven range shifts were probably made in a non-scientific context, for example, by fur trappers, fishers, or people with close economic or cultural connections to species. In fact, in some cases, such information (e.g., fur trade figures, traditional ecological knowledge) has been used to reconstruct species' historical ranges in large-scale biogeography studies $^{27-29}$. The first wave of scientific papers attributing range shifts to anthropogenic climate change occurred in the 1990s. These studies first observed changes in species distributions near range edges or shifts in thermal affinities among community members at static locations $28,30,31$, and eventually whole ranges shifting in response to climate change¹¹. Since then, range shifts have been observed across many taxa and environments^{15,19,22,26}, emerging as a widespread ecological response to climate change.

1.1 Measuring climate-driven range shifts

Estimating species distribution changes in response to climate change over time requires knowledge of where species currently are as well as where they once were, and ideally, climate variables over the relevant time period. In some exceptionally data-rich taxa and regions, this knowledge has been obtained by systematically monitoring occurrence and/or abundance of species with regular sampling at large spatial and temporal scales (e.g., National Oceanic and Atmospheric Administration fish trawls, used by $32,33$, Breeding Bird Surveys used by 34 , and the UK Butterfly Monitoring Scheme used by ³⁵).

When such high-resolution time series data are not available, species range movements can instead be inferred in other ways. When survey data are available in the present but not in the past, natural history records and museum collections can be used to reconstruct past species ranges to estimate distributional shifts $36,37$. When species range survey data are available in the past, conducting targeted re-surveys in the same regions years to decades later has been a relatively common method for assessing distribution changes^{38,39}. When data are available over broad temporal, but not spatial scales, assessing change in biogeographic communities at fixed locations^{30,40} can show relative abundance changes as "cross-sections" of shifting species ranges. Conversely, when data are available at broad spatial, but not temporal scales, comparing distributions of life stages of organisms at one point in time (e.g., seedlings versus adult trees⁴¹⁻⁴³) can reveal distributional changes in new versus old growth. Finally, new techniques such as telemetry⁴⁴ and citizen science observations can help to detect range shifts in not-previously-conventional ways $45,46$. These different methods have allowed rapid documentation of biogeographic change across regions and systems, together identifying range shifts as a relatively ubiquitous "fingerprint" of climate change $12,47$, but variation among methods has caused difficulty in synthesizing and comparing studies globally.

Box 2: Defining Ranges in Space

Range shift studies have taken a practical yet varied approach to defining range limits and centers despite an extensive literature on how species ranges could be defined $48-51$, and these variations could influence downstream comparisons. Methods commonly used to define species ranges in biogeography research include mapping based on gridded occurrence records, creating convex polygons or linear extents along a transect from irregular observation points, and more recently, using species distribution models to infer distributions based on environmental parameters when observational data are limited or disparate (reviewed in 52). Each of these methods are further affected by the definition of viable occurrences which are used as input data. For example,

depending on the resolution of the underlying data, the edge of a species range might be defined as the furthest straying individual at any life stage, the mean among multiple extreme individual observations, a high percentile among all observations, or the furthest known reproductively viable population (see figure). In addition, depending on the location and spatial extent of the study, the "range edges" or "range centers" defined within the study do not always represent the range edge or center of the global distribution. These varied definitions may cause problems when comparing findings between studies or even between sampling periods within studies, and can have a greater effect on range edges (that are defined by a lack of observations) compared to range centers 53 .

Practicality often calls for different definitions of range edges and centers for different data types. If a study relies on data from only a few sampling locations over a large timescale, a range edge defined as the furthest sampling location where the species is found may be appropriate. Alternatively, if a study uses high-resolution data such as regularly sampled grid cells over a continental scale, the authors may define the range edge as the average position of some percentile of occupied cells. In over 16,000 latitudinal range shift estimates included in the BioShifts database²², a diversity of definitions have been used for each range section, with some tendency of consensual definitions across studies (see figure). While consistency in the definition within taxa or studies can still provide estimates of shifts within studies (e.g. the location of the 95% percentile of occurrences moved polewards), differences in definitions among studies can limit the interoperability of range shift data when combining them, for example if some range definitions are more responsive to climate change than others.

1.2 Where range shifts have been detected

Range shifts have been assessed across continents and ecosystems, with reported observations highly skewed geographically towards Europe and North America^{15,54}. Indeed, among the over 26,000 latitudinal and elevational range shift estimates collated in the *BioShifts* database 22 , more studies appear to have been conducted and more species' ranges assessed per study in Europe and North America (Fig. 1a,b). We note, however, that the focus of this database on English-language papers could contribute to this bias^{55,56}. Observations are also biased taxonomically, with most range shift observations documented among terrestrial arthropods (including freshwater), vascular plants, and birds (Fig. 1a). In fact, of all latitudinal and elevational range shifts included in the database, 84% were from these three taxonomic groups and measured in Europe or North America (Fig. 1a), lending a disproportionate amount of our knowledge of range shift directions and speeds to a limited set of species, mostly in the Global North.

In addition to spatial and taxonomic bias, the range shift estimation effort has been unequal across ecological realms. The vast majority of estimates have been made in terrestrial systems, with fewer than 5% of all shifts in the BioShifts database in marine systems (Fig 1b, 2). Because marine and terrestrial species interact differently with their environments, it is likely important to generalize within each realm rather than across them. It is further plausible that freshwater species could show different responses to climate change than their terrestrial counterparts, but the terrestrial/freshwater distinction can be difficult to make as many terrestrial species rely on freshwater environments for certain life stages or behaviors, but are not necessarily bound to them for movement; for this reason, all species that live all or part of their life cycle in freshwater are considered as terrestrial in the BioShifts database and throughout this review.

Within species ranges, range shift estimates are made far more rarely at warm range edges compared to cold edges or range centers (11%, 41% ,and 48% of all shift estimates, respectively) (Fig 2). This bias might limit our understanding of mechanisms, as the processes driving range changes are expected to differ between sections of a range⁷. Assessments of range shifts encompassing both range edges within single studies are exceptionally rare – accounting for only 2% of range shift estimates in a recent continental review in Australia⁵⁷ – highlighting that the vast majority of our observations are capturing only glimpses of full species-range responses to climate change.

Figure 1: Geographical and Taxonomic Biases in Range Shift Detections.

a) Taxonomic and geographical breakdown of latitudinal and elevational range shift estimates in the BioShifts database²². Each tile represents 10 empirically-estimated range shifts of a given taxonomic group, always rounded up. b) Geographic bias of range shift studies in the BioShifts database. Each circle represents one range shift estimation study; circle sizes and inset numbers represent the number of species ranges assessed in each study, and color represents the type of shift estimated (marine latitudinal, terrestrial latitudinal, and terrestrial elevational shifts). Note that one species may have multiple appearances in a study if the study assesses multiple parts of the species range, or estimates shifts over multiple time periods, and a study may be represented in multiple continents if the study area spans across them.

1.3 General findings on range shifts

The direction of collated latitudinal and elevational range shift estimates in the BioShifts database are remarkably variable. Within the database, most range shift estimates (59% of all estimates) were in the direction expected based on local temperature change (here defined as the direction of isotherm shifts over time in the study region). Still, a high proportion of shift estimates (35%) are in the opposite direction as expected (opposite the direction of local isotherm shifts; Fig. 2), and a portion (6%) of species were not found to shift at all. These tallies of raw change in range position over time do not take into account uncertainty or distance of range shifts, and make null shifts methodologically unlikely to detect; nevertheless, even in this coarse tally, the proportion of estimates matching the direction of temperature change is greater than those opposing in all realms and range sections (Fig. 2, up vs. down arrows).

Along both latitudinal and elevational gradients, shifts at the cold edge and range center are more often directionally consistent with climate expectations than shifts at the warm edge (Fig.

2). While there are fewer shift estimates at warm edges in general (Fig 2), and extinction processes are more difficult to detect⁵⁸, the greater inconsistency with temperature shifts at warm edges could indicate that extirpation processes adhere less closely to temperature change than do range expansion processes (but see ⁵⁹), perhaps due to long life histories or compensatory processes^{60,61} among other hypotheses (see Section 2).

Considering the magnitude of range shifts reveals further consistencies with climate warming. Despite wide variation between individual range shift magnitudes, previous syntheses have found that range shifts have occurred faster in regions with higher rates of climate warming^{16,17,62}. Although there have been far fewer estimates of range shifts in marine species altogether and they are the least proportionally consistent with the directions expected from temperature change (Fig 2), a recent comparative study of range shift velocities that accounted for methodological variation and non-random taxonomic sampling found the velocity of range shifts to be higher in the ocean than on land, with marine species tracking isotherms more closely and moving more than five times faster than terrestrial $(5.9 \pm 0.9 \text{ km} \cdot \text{yr-1})$ for marine vs 1.1 \pm 1.0 km.yr-1 for terrestrial) on average in the poleward direction¹⁵. These patterns might be related to the tighter thermal niche 'filling' of marine organisms⁶³, the relative unavailability of microclimate refugia in marine habitats⁶⁴, the relative lack of human-created barriers in the ocean 65 , and the faster climate velocities in marine habitats than on land 8 , and suggest $^\circ$ mechanisms for slower or delayed range shifts on land (see Section 2).

Figure 2: Proportion of elevational, terrestrial latitudinal, and marine latitudinal range shifts consistent with climate expectations (i.e., direction of isotherm shifts) in the BioShifts database. Green and orange arrows signify range shift observations (raw changes in range positions over time) in agreement and opposition (respectively) to the direction of isotherm shifts in the study area, and tan segments show number of observations that estimated no range shift. Bars (excluding arrow

heads) are sized by the total number of detections in each group. Note that this graphic only shows agreement between biological and climate shifts, but not the direction of the shift; (e.g., in some cases, climate change yields a downslope expectation for elevational shifts, but are portrayed as part of the upward arrow here if the detected biological shift matches the isotherm shift).

Despite these generalized global patterns of species redistributions, a significant portion of range shift estimations do not match the rate or even direction expected from temperature shifts alone^{15,18,54,66}. The proportion of estimates consistent with the direction of temperature change across all realms is not overwhelmingly high (59%, Fig 2), and some range shift estimates vary in rate by an order of magnitude from their corresponding temperature shifts^{16,19}. While some of these inconsistencies could be due to methodological limitations or random variation 18 , others might be due to ecological processes that affect species ranges over and above temperature. These non-intuitive responses could be especially valuable to improve our understanding of factors that limit and alter species distributions.

Section 2 Why observations deviate from expectations

Several potential mechanisms have been proposed to explain why species ranges don't keep pace with changing isotherms, and why "naive" isotherm-tracking hypotheses may be insufficient for properly understanding species redistributions. These reasons can be summarized as (1) climate drivers other than (average) temperature, (2) non-climatic drivers of species ranges, and (3) low signal to noise in range shift detections. We unpack each of these below and review the literature on how these have been implicated in observed range edge responses.

Figure 3: Schematic of factors driving range shifts. (a) processes of species range shifts if temperature is the only driver, and (b) additional factors that may deviate/diverge range shifts from temperature-based expectations. In both panels, blue shaded regions represent the species range under historical climate conditions (T1), and red shaded regions represent the species range under future, post-warming conditions (T2). Dotted outlines represent the post-warming (T2) thermal potential range, or regions where temperature is suitable for the species. Inset plots in panel A show changes in thermal performance of populations fixed in space between T1 and T2. Features in panel B display various modifiers that can deviate/diverge range shifts from temperature-based expectations.

2.1 Climate drivers other than (average) temperature

There are multiple approaches to deriving expectations of the direction and velocity of species range shifts, and scale matters. In some cases, range shift expectations are derived from broad hypotheses based on global-scale temperature patterns, wherein species are expected to move towards higher latitudes, higher elevations, and deeper waters, without explicit consideration

of the actual climate gradient in the study region^{26,47}. These global-scale expectations can misrepresent local-scale changes experienced by organisms, as microclimates and climate refugia can misalign local gradients with global gradients; in other words, local isotherms may not be shifting polewards or upslope^{17,67,68}. Omission of local processes in favor of global-scale expectations can cause a higher proportion of detections to appear to defy expectations 54 . However, methods for including micro-scale changes are being developed $69,70$, and data to do so are increasingly available^{67,71,72}. These advances can even allow for *a posteriori* reassessment of range shift studies⁷³. For example, when local isotherm velocities were calculated for range shifts in the BioShifts database 15 following methods from 8 , about 2% of elevational study areas and 5% of latitudinal study areas had isotherm velocities in directions opposite of global gradients (that is, downslope or equatorwards). If species perfectly tracked temperature changes, situations like these would manifest as expectation-contrary shifts if only global gradients were used to form expectations.

Even at relevant spatial scales, temperature is not the only limiting climatic variable of species ranges. Different climate variables can show contrasting patterns of change through time⁷⁴, so if a species' most limiting abiotic dimension is not temperature, that species can show shifts contrary to temperature-based expectations. For example, some species shifts have more closely tracked changes in precipitation or water balance than temperature, even when those environmental changes are directionally misaligned with temperature shifts^{75,76}. In other cases, species range shifts have been more closely correlated with niche-tracking (over multiple abiotic dimensions) than isotherm-tracking alone^{$77,78$}. Still, these non-temperature variables are rarely incorporated when forming expectations of climate-driven range shifts, and likely explain a portion of the temperature-contrary shifts seen in global estimates. Moreover, constraints imposed by abiotic gradients that are not changing with climate change can be important⁷⁹. For example, day length may constrain photosynthetic processes of plants and corals from shifting across latitudes $80,81$, and atmospheric pressure and oxygen availability may constrain animals from shifting up elevations^{79,82}. Interactions between variables with warming might also lead to expectation-contrary shifts, such as the link between temperature and oxygen demand driving oxygen-supply limitations downslope on warming mountains⁸² or offshore in marine coastal systems⁸³.

Even in cases where temperature is the primary driver of species' responses and the variable forming expectations, the chosen metric of temperature (e.g., annual mean, annual max, number of days over temperature threshold) may not best represent the limiting factor for the species. Temperature extremes, growing degree days, duration of summer, or temperaturerelated landscape variables such as spring snow cover might variably be more direct drivers of ecological responses to climate change $84-87$. Any of these variables can affect species differently throughout their life histories, rendering average temperature a coarse predictor of species responses. Temperature anomalies that are not resolutely captured by average trends can also acutely affect species over and above long-term changes. For instance, marine heatwaves over the last decade have exacerbated poleward range shifts in shallow reef species in addition to gradual change through time⁸⁸. Long-term temperature averages might therefore serve as a useful proxy to capture species responses to many correlated temperature variables in largescale studies, but might not drive responses of individual species.

2.2 Non-climatic drivers of species ranges

Species can be excluded from portions of their climatically suitable niches for many reasons, and the extent to which species' ranges are in equilibrium with their environments can affect the predictability of their responses to climate change^{63,89}. Limitations indirectly or fully unrelated to climate change can pose constraints on species' realized ranges (Fig 3), and result in ranges showing inertia or climate-change-contrary shifts. Here we organize these nonclimatic range-limiting factors into four categories – habitat limitations, dispersal constraints, biotic interactions, and contemporary evolution – and briefly review evidence about how each alters leading or trailing range shift rates against temperature-based expectations.

2.2.1 Habitat Limitations

Habitat limitations caused by fragmentation, geographic barriers, and human impacts can either prevent or facilitate range shifts. Geographic barriers such as mountain ranges, valleys, and coastlines can present impasses for range-shifting species, regardless of climatic suitability beyond these barriers³⁸. For species that are bound to geographically-constrained habitats such as rivers, coastlines, or high-altitudes (i.e. sky islands), landscape connectivity can be much more influential than climate in determining distribution ranges⁹⁰, and can result in overall range reductions when ranges can contract at one end but cannot expand at the other⁵⁹. When habitat availability does not align with latitudes, elevations, or climatic gradients over which range shifts are expected, species may be unable to keep pace with shifting isotherms (e.g., fish in an east-to-west river can't migrate north) 91 and thus might present as examples where species exhibit little or no range shifts in response to climate change.

Even when a habitat generally extends in the direction of climate velocity, habitat fragmentation and land use change can inhibit range expansions at the cold edge or amplify range contractions at the warm edge⁹²⁻⁹⁴ (Fig. 3b). Human modifications like forest clearing can change the overall permeability of landscapes, which can slow or stop range shifts of forest species through cleared areas^{95,96}; landscapes with higher conductance are most quickly shifted across⁹⁷. Alternatively, conserving habitat through establishing protected areas can facilitate species range shifts compared to surrounding areas⁹⁸ (Fig. 3b). Of course, in cases where

species specialize in such human-modified landscapes (e.g., farmland-associated birds or mussels on man-made offshore structures), human modifications might represent increased connectivity and actually facilitate range shifts^{99,100}. Some projections of climate-driven range shifts account for geographic barriers by including landscape connectivity in models or projecting range-shift corridors^{101,102}, but such considerations are normally unaddressed in range shift detections and may present as unexplained lags or advances compared to climate tracking.

2.2.2 Dispersal Constraints

Climatically suitable areas may exist unobstructed by physical barriers, but remain inaccessible to species that lack a mode of dispersal to reach them, thus slowing range shifts from temperature-based expectations. Dispersal of sessile or slow-moving species such as benthic marine invertebrates and wind-dispersing plants relies on external directional forces (ocean currents or prevailing wind directions), and these forces may be misaligned with the direction of isotherm shifts (Fig. 3b). These misaligned dispersal processes can impede species from tracking changing environments, particularly at the leading edge $103,104$. Animal-assisted dispersal can also impede range shifts of seed-bearing plants as animal vectors do not always disperse in the direction of climate shifts^{105,106}. Among actively dispersing organisms, like most terrestrial vertebrates, dispersal abilities can vastly differ, in some cases limiting species' ability to track environmental change. In addition, when there are few dispersers, Allee effects and inbreeding depression can limit the success of early arrivals, leading to lags between species responses and climate shifts 107 .

The role of dispersal ability is typically assessed using various proximal traits such as body size and reproductive mode¹⁰⁸⁻¹¹¹. Although strong relationships between these traits and range shift rates are not common, these expectation-contrary results could be a greater reflection of inadequate estimates of dispersal ability (in which proxies can perform poorly) than a true absence of effect 112 . For more on dispersal and species traits, see 110,112 .

2.2.3 Biotic interactions

Biotic interactions can affect a species' ability to redistribute under climate change and cause lags or unexpected range shifts relative to climate expectations. Species interactions are expected to affect leading and trailing range edges differently according to whether they are positive (e.g., food, mutualism, commensalism) or antagonistic (e.g., predation, parasitism, ${\sf connection})^{\rm 113}.$

Positive interactions can result in reduced or delayed range shifts at either range edge. At the leading edge, a species may be unable to shift in response to climate change if its distribution is bound through an obligate positive interaction with another species (e.g. a specific habitat,

host, or food resource species) that is less able to expand in pace with isotherms. Such limitations have been demonstrated in butterflies shifting up mountain slopes being limited by distributions of host plants which are shifting more slowly^{114,115}. At the trailing edge, positive interactions such as fungal mutualists on plant roots can broaden the environmental tolerance of the interacting species such that populations with mutualists contract more slowly than populations without in response to environmental change¹¹⁶⁻¹¹⁸. Positive interactions can also manifest as a direct alteration of climate conditions, such as when desert plants create microclimates for associated species; these can appear to alter the rate of climate responses, particularly when expectations are built on macro-level climate patterns that don't capture the microclimate conditions experienced by organisms (see Section 2.1).

Antagonistic interactions can lead to slower range shifts than expected. If a species has a trailing edge limited by antagonistic interactions (but not by climate), the trailing edge shift of the focal species could be driven by the speed of the leading edge shift of the antagonistic species, which, depending on climate-tracking ability of the antagonistic species and the timecourse of the interaction, might be slower than climate^{119,120}. The reverse is also possible, in which a species' leading edge is limited by an antagonistic species' trailing edge that is not (or is slowly) responding to climate change. Such is the case in intertidal habitats where mussels and barnacles that are shifting towards deeper, cooler waters, have their leading edge (the deeper edge) limited by presence of sea star predators 121 . But this scenario requires a nontemperature reason (in the case of sea stars, high-tide desiccation stress) for why the antagonistic species' trailing edge (the shallower edge) is fixed and not responding to warming.

In some cases, climate-consistent shifts of one species could lead to climate-contrary shifts of an interacting species. For example, if two competing species had overlapping ranges, but only one was sensitive and responded to temperature change, warming could lead to an expected range shift in one species, and an increase in abundance near the warm range edge in the other species where the competitor vacates 122 . This scenario, and many of those previously mentioned, hinges on differences in the climate responsiveness of interacting species, the frequency of which is plausibly high but generally unknown. Methods to model distributions of interacting species are becoming more common and might be able to disentangle some of these nuances. For more on multispecies modeling, see $^{123-125}$.

While theory supports a number of mechanisms by which biotic interactions alter range shift rates compared to climate expectations¹²⁶, they are difficult to observe. In the above examples, species-interaction-driven range shifts (or lack thereof) were identified through careful natural history observations and/or directed research. However, most large-scale range shift studies assess many species' ranges at once and typically lack context of species-specific climate

tolerances or community interaction webs. Without independent knowledge of species' interactions, the previously discussed examples would have presented as species responding in unexpected ways to climate change with the underlying mechanisms unrecognized. Thus, it is likely that unidentified biotic mechanisms underlie a portion of the expectation-contrary shifts that have been documented.

2.2.4 Contemporary evolution

Contemporary evolution can affect the rate of species range shifts in response to climate change 127 , although a full summary of the theoretical directions and mechanisms is beyond the scope of this review. Briefly, evolution of dispersal ability can increase the rate of range expansions if previously unsuitable areas become suitable beyond the species range^{128,129}, just as evolution of any trait that alters the role of range-limiting factors (desiccation resistance, species interactions, heat tolerance) can affect the response of ranges away from expectations^{130,131}. The challenge to invoking trait or niche evolution during range shifts is explaining why these traits would evolve more readily during a range shift when they could have been under selection in previous climatically-stable periods¹³², i.e. ultimately understanding the limits to niche width evolution at all range edges. For more on evolution and range shifts, see 127,133,134 .

2.3 Low signal to noise in range shift detections

Methodological differences overall can greatly influence estimates of range shifts from empirical data. In fact, two recent syntheses of species range shifts made similar conclusions that methodological factors are better predictors of range shift rates than ecological traits or climate variables^{15,18}. Because range shift detections often rely on sparse or opportunistically collected data, researchers often have little choice in methodology for individual studies. This range of methodologies adds noise that may impede our ability to detect meaningful signals of change or draw synthetic conclusions between systems 18 . Since species ranges are naturally variable from vear to year⁷, and are themselves observed imperfectly 37,53,58 , low sampling frequency or spatial coverage can result in estimates that are misrepresentative of or even contrary to long-term trends³².

High noise in range shift estimates can be expected when the temporal extent of a study is small (i.e., a short total duration of observations). Even if the pace of climate change has been fast within this time frame, life-history constraints and stochasticity likely affect range edges in the short term, thus time can be required for robust changes to be detected at range edges. Indeed, both latitudinal and elevational shifts of range edges from the BioShifts database show patterns of high variability around the temperature-based expectation in short-duration studies; that variability diminishes towards zero (matching expectations) as study duration

increases (Fig. 4), suggesting that, at least among longer term studies, temperature is a main direct or indirect driver of range shifts. This trend is promising, but unfortunate, as reconstructing high-resolution biogeography information for the past is difficult. Nevertheless, it lends support to the value of reassessing documented range shifts in cases where sampling can be continued after an initial study is completed.

Figure 4. Lag between isotherm shifts and range shifts across study durations in the BioShifts database. Lag is calculated as shift velocity - isotherm velocity, where a perfectly isotherm-tracking shift would have a lag of zero. Purple dots signify range shifts estimates of trailing edges, green of leading edges for both elevational and latitudinal range shifts. Dashed lines show the 5th and 95th quantile regressions of all points in each plot.

Related to the role of time, estimates made from too few observations can carry both process and sampling error, affecting their accuracy. Range edges are particularly prone to low-accuracy shift estimates based on the likelihood of error of both types. Process error can greatly influence range edge identifications when stochastic events and idiosyncratic biological processes vary their positions from year to year (e.g., higher than normal wind dispersal in some years). Moreover, since edges normally contain lower population densities of focal species¹³⁵, position estimates can be substantially affected by sampling effort, such that low sampling can add noise to already variable range edges 53 . While studies with many observations of range edge positions through time may be able to cut through this error to detect biological signal, those with sparse data are likely confounded by these issues to some extent.

Statistical limitations can further complicate range shift detections, and are often unaddressed in range shift detection studies 136 . For example, when range edge estimates are expected to be variable across observations (either because of natural range edge variability or observer error), the "regression towards the mean" effect can lead to an apparent range shift in which an estimated range edge is shifted towards the location of the true range edge (the expected sample mean), even in the absence of a true biological change^{137,138}. In addition, because

sampling of species range position is done over a finite sampling domain (e.g. transects up mountains or across latitudes), range shifts that surpass the domain will necessarily be underestimated. Unless removed from the study, these "observation domain crossers" will lead to predictable correlations between the starting range edge relative to the domain edge and the magnitude of the detected shift¹³⁷. Despite these potential sources of error, fewer than 2% of range shift detections are tested against null models¹³⁶, resulting in an inability to distinguish whether range shift detections differ from range shifts due to chance alone.

Detecting range edge shifts based on the furthest occurrence of a species at only two points in time without comparison to a null model is one situation in which noise might particularly outweigh biological signal. Yet due to understandable limitations in data availability, scenarios like this are quite common; of the latitudinal range shifts in the BioShifts database, 10% of cold edge and 20% of warm edge estimates used furthest-occurrence detections to establish range edge positions (see Box 2), and in two recent syntheses of range shift estimates, 38 and 85% of range shift estimates (cold and warm edge, respectively) were made using two time-point $data^{18,22}$. Various approaches can be taken to reduce the influence of these sources of noise when comparing range positions between few time periods or with sparse data $37,139$. Some common strategies include aggregating observations across multiple years to reduce the influence of natural variation and observer bias, accounting for sampling effort and detectability in occupancy models, and aggregating error-prone shifts across multiple species to form an understanding of community shift rate among noisy individual estimates. Future work should consider estimating error in range edge detections, including meta-analytic methods and null models that account for variation in precision among studies.

Section 3 Impacts on natural and human systems

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Range shifts and their cascading effects will lead to both positive and negative outcomes for species, ecosystems, and people. Range-shifting towards suitable habitats is, in itself, an adaptation that allows for persistence in the face of ongoing climate change. Therefore, climate-driven leading edge shifts have broadly positive effects for individual species and for maintaining biodiversity at large. Still, rapid changes in species ranges have led to corresponding needs for adaptation in natural and human systems that were established under more stable conditions. Among social-ecological coupled systems, range shifts are expected to have greater impact on those that lack the financial or political capacity to quickly adapt 20,21 . creating possible equity imbalances that should be addressed in future research.

Species shifting into new areas at their leading edges occupy a unique position in science, policy, and management. Although climate-induced shifts of species' "native" ranges

fundamentally differ from the spread of invasive alien species, existing policies are not consistently or decidedly equipped to manage them. Due to the vast potential impacts of newly arriving species, understanding and predicting range shifts has been increasingly prioritized by governmental research agencies $140-142$, and their perception by resource managers has varied from new exploitable resources to dangerous invaders $143,144$. Here we provide a categorical overview and an update of major cascading effects of range-shifting species on ecosystems and people, focussing on impacts of range shifts that have already occurred rather than future projections, but direct interested readers to 13 for a more in-depth review.

3.1: Impacts on Natural Ecosystems

Redistributions of species can lead to changes in community composition and dynamics, or otherwise alter natural ecosystems, and leading edge shifts can have effects akin to those caused by non-native biological invasions $14,120$. When range shifts of multiple species occur in a region, the reshuffling can result in cold-adapted species being replaced by warm-adapted ones, referred to as 'thermophilization', thus altering the historical structure and function of the community¹⁴⁵⁻¹⁴⁹. Despite changes in distributions of individual species, the ecosystem as a whole might maintain function if species lost due to range shifts are functionally replaced by species entering, or if species change their functional contributions as communities change 150 . Range shifts can also lead to loss of specialist species in favor of generalists, potentially reducing biodiversity at regional scales^{151,152}. The magnitude and rate of thermophilization can vary between taxonomic groups, biogeographic contexts, or species traits related to persistence and mobility, but changes in thermal affinities of communities tend to lag behind climate warming in general^{40,153}, possibly leading to changes in the stability of communities as they become increasingly out of sync with their environments.

Some species range shifts can have disproportionate impacts on natural ecosystems, such as changes in habitat-forming foundation species. For example, the barren-forming urchin, *Centrostephanus rodgersii*, has exhibited a poleward expansion along the south-eastern coast of Australia in recent decades, leading to overgrazing of large areas of kelp forests and removing key habitat for kelp-associated species including those comprising major fisheries exports of the region¹⁵⁴⁻¹⁵⁶. Alternatively, global mangrove species have exhibited poleward expansions in recent decades¹⁵⁷ that could disproportionately change the composition of communities in areas in which they have established, given that mangroves provide essential habitat for a wide range of species¹⁵⁸. Shifts of species that shape their environments can open niche spaces and promote shifts of other species, resulting in increased species richness or phase changes to new ecological communities 149,159 .

3.2 Impacts on Humans

Climate-mediated shifts in species distributions can have direct consequences on human communities, inclusive of economic and cultural systems with which they are connected. Shifting ranges of food species can lead to positive or negative outcomes for human communities, and require adaptation in order to begin harvesting and managing newly-arrived species, or transition away from reliance on outward-bound species. For example, distributional shifts of commercially important fisheries species have created opportunities to open commercial and recreational fisheries as new species arrive in some cases 160 , but reduced fishing employment¹⁶¹, and required increased travel for harvesting as target species shift in others¹⁶². Similarly, crop species have shifted ranges in recent decades, leading to economic opportunities in places that have become favorable, and necessity for adaptation in places that have become unfavorable¹⁶³; crop pests and pathogens have also shifted, changing the spatial distribution of some food security threats 164 .

The effects of species redistributions on human cultural systems can be considerable. For indigenous communities, species range shifts can threaten abundance of culturally important resources¹⁶⁰, access to traditional foods¹⁶⁵, and traditional knowledge systems built around natural ecosystems^{20,165}. Range shifts can also occur in species that are emblematic of the places where they were historically found¹⁶⁶, potentially altering societal connections with local ecosystems.

Species range shifts can also present threats for human health. Documented range shifts of disease-vector species due to climate change have resulted in spatial changes to health risks for human communities. Malaria-vector mosquitoes in Africa have shifted upslope by about 6.5 m per year and poleward by about 4.7 km per year since the late $1800s^{167}$. Similar shifts of mosquitos and other disease-bearing species such as ticks and bats have been documented around the world¹⁶⁸⁻¹⁷⁰ and are expected to continue in the future^{171,172}. Ciguatera poisoning, caused by eating certain types of reef fish found typically in warm tropical water, has seen a geographical expansion of poisoning incidents since 2000, as warm-water fish shift polewards along the east Australian coast¹⁷³. Public-health-relevant range shifts have direct implications, introducing regions and communities to risks for which they are historically not prepared and requiring preparation from healthcare systems to manage them. Meanwhile, range shifts of disease-bearing species could – in theory – relieve human populations from diseases that have historically been prevalent as trailing edges of their ranges contract upslope or away from the equator. Some such cases have been observed 174 , and others are predicted in coming decades^{171,175}. In addition to direct range shifts of diseases and their established vectors, novel species interactions from range shifts are one factor leading to an increase in risk of zoonotic

spillover of pandemic-causing viruses, and a majority of these novel interactions to come in the next century might have already occurred 176 .

3.3 Impacts on resource management and conservation

Most conservation decision-making processes have been built around assumptions of relative stationarity – that is, assuming current ranges will persist into the future¹⁴³. The global redistribution of species in response to climate change breaks this assumption, and management actions (including land acquisition, habitat management, and direct population management) will need to incorporate projected impacts of range shifts in order to achieve optimal or workable outcomes. Protected areas, for example, have been useful tools for preserving biodiversity, abundance, and assemblage structure within them¹⁷⁷, but risk losing their effectiveness (depending on their stated goals) when species that inhabit them shift away and new species enter their boundaries^{178,179}. In anticipation of species range shifts, conservation managers have needed to re-evaluate their goals and adapt their strategies in order to preserve biodiversity into the future.

One approach to management in light of anticipated range shifts is to supplement permanent protective measures with temporary efforts as needed, effectively increasing connectivity to sufficiently protect range-shifting species¹⁸⁰. This strategy has been implemented in marine environments by adding temporary protective measures to increase connectivity between permanent sites or to protect new marginal populations¹⁸¹, adapting protected areas to shift over time in accordance with gradual shifting of species¹⁸², designing protected areas as stepping stones across climatic gradients¹⁸³, and streaming low-latency ("real-time") data to project species distributions and guide harvesting and management $184,185$. While less common in terrestrial settings, similar dynamic conservation goals have been achieved by implementing short-term conservation areas for migratory species through "rentals" of private land 186,187 . While the stated fundamental objectives of most conservation programs appear to be at odds with time-limited protected status, combining temporary conservation areas with traditional permanent protected areas might help to maximize conservation benefit with limited funds.

A related strategy is targeting establishment of new protective measures within "climate corridors" where species range shifts are likely to occur. For example, The Nature Conservancy (USA) recently implemented the Cumberland Forest Project¹⁸⁸ $-$ one of its largest land acquisition projects to date — specifically intended to protect key "climate escape routes" for range-shifting species. Priority for range-shift corridors has also already been integrated in the biodiversity guidelines of the US Department of Agriculture's National Agroforestry Center 189 .

Climate-driven range shifts have also spurred novel strategies in biodiversity conservation. In light of shifting climatic suitability, assisted migration has emerged as a conservation strategy wherein individuals are transplanted either within or outside of their current geographic range with the intention of situating them where they are more climatically adapted¹⁹⁰. Although controversial¹⁹¹, some assisted range migration efforts have already begun¹⁹²⁻¹⁹⁵. To facilitate persistence at the warm edge, conservation efforts that protect genotypes pre-adapted to future conditions are particularly effective¹⁹⁶. These efforts facilitate evolutionary rescue, a process likely to be particularly important to maintain biodiversity in the tropics¹⁹⁷, whilst also bringing to light ethical considerations regarding where and for which species such efforts should be applied.

Given their importance to human systems, range shift projections have highlighted the need for transboundary management strategies as species shift across political or jurisdictional borders¹⁹⁸. Transboundary conflicts have already arisen when commercially important species have shifted to neighboring jurisdictions¹⁹⁹, and have been addressed as a priority in international resource management in the Anthropocene^{143,200}. Needs for proper transboundary strategies will only become greater, given that models predict vast swaths of future habitat for many species to be in countries where those species are not currently found²⁰¹⁻²⁰³, and many of these changes will disproportionately impact countries with high reliance on resource species and low capacity to adapt²⁰. Without proactive management and transboundary agreements in place, species shifting from one historical jurisdiction to another may result in an incentivized depletion of the resource by the country the species is shifting away from that stands to lose future access.

Box 3: Human Range Shifts

Shifting isotherms across Earth's surface are changing climate conditions experienced by humans, too. Despite our great powers of niche construction and climate buffering compared to many other species, human distribution for centuries has clustered in a surprisingly narrow envelope of climatic conditions²⁰⁴. Anthropogenic climate change is shifting this climatic envelope away from existing human population centers, especially those in tropical climates, and creating novel challenges, opportunities, and ethical considerations regarding "range shifts" of our own occupancy patterns and activities in a warming world 205 .

Range shifts of human activities due to climate warming have already begun, both at contracting warm margins and expanding cold margins. Warming temperatures over recent decades have already led to poleward and upslope shifts of agricultural activities, including latitudinal shifts of "breadbasket" and "rice bowl" crops across continents¹⁶³ and upslope shifts of mountain crops like coffee^{206,207}. In the ocean, historical patterns of human ocean use have expanded concurrently with warming, with fishing fleets shifting polewards by almost 1

degree latitude per year²⁰⁸ and trans-Arctic shipping routes having greatly expanded in navigability since the 1980s²⁰⁹. Warming conditions have also changed spatial patterns of scientific pursuit, resource extraction, and even recreation and tourism across latitude and elevation^{210–216}. These changes to human behaviors may already be having downstream effects on natural environments, adding additional pressures to systems already facing stress from changing climates.

Whereas climate-driven range shifts of human activities have been documented, direct climate-related changes to human distributions are more difficult to attribute and predict, because many factors, including politics, wealth distribution, infrastructure, and cultural ties influence people's decisions and ability to migrate. Shifting human climate envelopes underscore the need for robust adaptation strategies when climate-tracking migration is not possible; recent models predict large areas of extreme climates, even surpassing human physiological tolerance thresholds, in the near future^{204,217}. Moreover, these trends call attention to the global ethical considerations of shifting climate regimes, as many regions at the warm margin of the shifting human niche are those contributing the least to global climate change 205 .

Section 4 Summary and Future Perspectives

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In the above sections, we summarized evidence that species' range shifts are a ubiquitous response to climate warming, including where and how they have been documented, why they do not always match expectations, and the impacts they have had on nature and society. Despite considerable advances in understanding mechanisms by which species redistribute in response to climate change, there is vast room for progres in the field. Here we outline three key priority areas to advance how we understand and predict biodiversity redistribution and improve our ability to make sound management decisions about species 'on the move' and their associated impacts.

4.1 Increased effort in monitoring and harnessing of data

Existing monitoring programs have contributed greatly to observations of climate-driven range shifts, but are often limited to particular ecosystems, geographical areas, or taxa. While monitoring programs can be less tempting to prioritize for funding agencies because benefits largely accrue over longer time frames, increased support for these efforts will lead to greater ability to observe, attribute, and anticipate species responses to climate change²¹⁸. Increased monitoring could be particularly valuable in systems or locations where current data are limited, or in which species ranges are already predicted to shift, to test and improve model prediction skill $62,219$. Furthermore, efforts to inter-connect disparate existing monitoring programs so they can be integrated, and include comparable and pertinent indicators of

biodiversity change (e.g., Essential Biodiversity Variables²²⁰), can help us to up-scale existing monitoring investments for inference at larger scales, such as whole-range biogeographical shifts 221 .

When systematically surveyed "gold standard" datasets aren't available, range shift research could benefit from increased use of publicly available citizen- and community science data. Such volunteer-driven data are collected around the world and already account for over half of biodiversity occurrence records in the Global Biodiversity Information Facility (GBIF) database²²². These records have already shown applications in biogeography research^{57,223-225}. with ongoing citizen science monitoring programs having contributed greatly to our observations of range shifts within systems (e.g., Christmas Bird Count, used in $34,226-228$). Outstanding barriers to further integrating citizen science data into climate change biogeography research include the strong spatial and temporal biases of data collections, the statistical measures needed to account for unstructured data collection (but see 229), and the validation of data from non-expert contributors (although robust post-hoc verification measures are possible, e.g., 223). Further research into how to control for these variables to best apply public data sources will open research pathways in the field.

4.2 Understanding mechanisms to increase generalizability

Synthesis of range shift studies has offered valuable insight into general trends and variability in observed species range shifts, but better understanding of the sources of variability between studies and the influence of data biases are needed in order to generalize. Identifying organism or environment-specific factors that make range shifts likely will improve our ability to anticipate them in cases where observational data is limited.

Assessing the influence of species traits on range shifts, for example, should help to generalize the types of species most likely to shift at their leading and trailing edges. Intrinsic factors such as a species' ability to move, its level of generalism, or its lifespan could influence its ability to shift in response to climate change, and theory suggests varying ways by which they should explain range shift rates $^{10,112,230-232}$. Yet evidence to date suggests fairly weak or variable explanatory power of traits in relation to observed range shift patterns (reviewed in 112), suggesting the methods and perhaps quality of data used to assess their influence have been lacking¹¹². Further research into intrinsic sources of variation could benefit from simulations and theoretical models to understand the influence that traits might have on range shifts within biological communities²³³, combined with data synthesis to test theories against empirical data.

Another promising direction is generalizing how the spatial pathways of species shifts interact with properties of the landscape. For instance, global climate velocities can be used to identify regions where range shifts might be expected²³⁴, global ocean and wind currents can identify areas where these forces will facilitate or impede shifts of passively-dispersing species^{103,104}, and landscape *quality* maps can estimate potential routes through which range shifts are more likely to occur²³⁵. Such area-based approaches could aid in management efforts to facilitate climate-driven range shifts even in systems where species specific data are unavailable.

Once mechanisms have been identified that can explain variation in species' range shifts, a promising avenue will be to integrated these into mechanistic or process-based models that simulate the eco-evolutionary response of species in face of climate change²³⁶⁻²³⁸. Integrating variables such as dispersal, evolutionary potential, or demographic processes into models can improve model fits and extrapolations over correlative approaches^{239,240}, and have the advantage of flexibility as key parameters can be estimated or imputed when species-specific data is unavailable²⁴¹. Reliance on underlying mechanisms (such as species traits or landscape habitats) makes models of this type transferable to systems where statistical models may be limited by data scarcity. Moreover, mechanistic models can be used in simulation experiments to better identify specific drivers of range shifts with other factors being held constant. Processbased simulations can control for intrinsic and extrinsic factors affecting range shifts, as well as natural variation and detection error to assess the effects of potential drivers of range shifts^{197,242}, and can even be used to assess effects of management interventions on species distributions²⁴³. The theoretical relationships that arise from simulations experiments can then inform needs to data collection that can circularly aid in improving *in situ* range shift observations.

4.3 Improving predictions at usable scales

While observing and detecting impacts of climate change is a necessary step in understanding processes, projections of range shifts into the future can inform mediation and management strategies to avoid or adapt to severe ecological impacts of climate change. In range shift research, projections are often made in the long-term (e.g., year 2100) using tools such as species distribution models or other correlative models of range position based on environmental variables; predicting range shifts in near-term, management-relevant timescales is much less common. This dearth of near-term predictions is one barrier to implementation of effective management strategies for species redistributions.

Near-term forecasting has grown in popularity and priority in recent years across subfields in ecology^{244,245}. In some data-rich systems with direct applications, near-term forecasts of species distributions have already been implemented to support dynamic management strategies $^{246-250}$. However, in most systems, such efforts have not been implemented. We suggest predictive efforts focus on systems in which predictions can occur in tandem with ongoing monitoring

systems to ensure that near-term predictions can be tested, falsified/validated, and improved as monitoring continues²⁴⁴. Focus on statistical tools that can accurately characterize and project uncertainty, as well as those that can integrate multiple processes from environmental suitability to landscape connectivity (see previous section) will be an important step in improving the predictive power of the field.

Conclusion

Despite significant strides in our understanding of climate-driven species redistributions in recent decades, much remains to be understood and discovered in the field. Since the field began, observations of climate-driven range shifts have been made around the world 22 . revealing both clear trends in observed responses, and high variation due to observational, methodological and biological factors^{15,18,136}. Given the recent growth of the field of climate change biogeography and the urgency to predict climate change responses globally, it seems projections, models, and theory have outpaced empirical observations in recent years, and our collective knowledge and abilities to predict stand to benefit from further efforts to augment empirical data for testing and validation. Expanding research efforts in the directions set out above will help us identify range shifts in the present, anticipate range shifts in the future, and use this knowledge to inform strategies for preserving biodiversity through climate change.

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Bibliography

- 1. Grinnell, J. Field Tests of Theories Concerning Distributional Control. *Am. Nat.* **51**, 115–128 (1917).
- 2. Colwell, R. K. & Rangel, T. F. Hutchinson's duality: The once and future niche. *Proc. Natl. Acad. Sci.* **106**, 19651–19658 (2009).
- 3. Soberón, J. Grinnellian and Eltonian niches and geographic distributions of species. *Ecol. Lett.* **10**, 1115–1123 (2007).
- 4. Soberón, J. & Nakamura, M. Niches and distributional areas: Concepts, methods, and assumptions. *Proc. Natl. Acad. Sci.* **106**, 19644–19650 (2009).
- 5. Brown, J. H., Stevens, G. C. & Kaufman, D. M. The Geographic Range: Size, Shape, Boundaries, and Internal Structure. *Annu. Rev. Ecol. Syst.* **27**, 597–623 (1996).
- 6. Anderson, B. j *et al.* Dynamics of range margins for metapopulations under climate change. *Proc. R. Soc. B Biol. Sci.* **276**, 1415–1420 (2009).
- 7. Lenoir, J. & Svenning, J.-C. Latitudinal and Elevational Range Shifts under Contemporary Climate Change. in *Encyclopedia of Biodiversity* 599–611 (Elsevier, 2013). doi:10.1016/B978-0-12-384719-5.00375-0.
- 8. Burrows, M. T. *et al.* The Pace of Shifting Climate in Marine and Terrestrial Ecosystems. *Science* **334**, 652–655 (2011).
- 9. Ling, S. D., Johnson, C. R., Ridgway, K., Hobday, A. J. & Haddon, M. Climate-driven range extension of a sea urchin: inferring future trends by analysis of recent population dynamics. *Glob. Change Biol.* **15**, 719–731 (2009).
- 10. Bates, A. E. *et al.* Defining and observing stages of climate-mediated range shifts in marine

systems. *Glob. Environ. Change* **26**, 27–38 (2014).

- 11. Parmesan, C. Climate and species' range. *Nature* **382**, 765–766 (1996).
- 12. Parmesan, C. & Yohe, G. A globally coherent fingerprint of climate change impacts across natural systems. *Nature* **421**, 37–42 (2003).
- 13. Pecl, G. T. *et al.* Biodiversity redistribution under climate change: Impacts on ecosystems and human well-being. *Science* **355**, eaai9214 (2017).
- 14. Sorte, C. J. B., Williams, S. L. & Carlton, J. T. Marine range shifts and species introductions: comparative spread rates and community impacts: Range shifts and non-native species introductions. *Glob. Ecol. Biogeogr.* **19**, 303–316 (2010).
- 15. Lenoir, J. *et al.* Species better track climate warming in the oceans than on land. *Nat. Ecol. Evol.* **4**, 1044–1059 (2020).
- 16. Chen, I.-C., Hill, J. K., Ohlemüller, R., Roy, D. B. & Thomas, C. D. Rapid Range Shifts of Species Associated with High Levels of Climate Warming. *Science* **333**, 1024–1026 (2011).
- 17. Pinsky, M. L., Worm, B., Fogarty, M. J., Sarmiento, J. L. & Levin, S. A. Marine Taxa Track Local Climate Velocities. *Science* **341**, 1239–1242 (2013).
- 18. Brown, C. J. *et al.* Ecological and methodological drivers of species' distribution and phenology responses to climate change. *Glob. Change Biol.* **22**, 1548–1560 (2016).
- 19. Poloczanska, E. S. *et al.* Global imprint of climate change on marine life. *Nat. Clim. Change* **3**, 919–925 (2013).
- 20. Melbourne-Thomas, J. *et al.* Poleward bound: adapting to climate-driven species redistribution. *Rev. Fish Biol. Fish.* **32**, 231–251 (2022).
- 21. Alexander, K. A. *et al.* Equity of our future oceans: practices and outcomes in marine

science research. *Rev. Fish Biol. Fish.* **32**, 297–311 (2022).

- 22. Comte, L. *et al.* BioShifts: a global geodatabase of climate-induced species redistribution over land and sea. 5238495 Bytes (2020) doi:10.6084/M9.FIGSHARE.7413365.V1.
- 23. Humboldt, A. von & Bonpland, A. *Essai sur la géographie des plantes :accompagné d'un tableau physique des régions équinoxiales, fondé sur des mesures exécutées, depuis le dixième degré de latitude boréale jusqu'au dixième degré de latitude australe, pendant les années 1799, 1800, 1801, 1802 et 1803 /par Al. de Humdboldt et A. Bonpland ; rédigée par Al. de Humboldt.* (Chez Levrault, Schoell et compagnie, libraires, 1805). doi:10.5962/bhl.title.9309.
- 24. Parmesan, C. Ecological and Evolutionary Responses to Recent Climate Change. *Annu. Rev. Ecol. Evol. Syst.* **37**, 637–669 (2006).
- 25. Odum, E. P. The Concept of the Biome as Applied to the Distribution of North American Birds. *Wilson Bull.* **57**, 191–201 (1945).
- 26. Thomas, C. D. Climate, climate change and range boundaries. *Divers. Distrib.* **16**, 488–495 (2010).
- 27. Gonzalez, P. Desertification and a shift of forest species in the West African Sahel. *Clim. Res.* **17**, 217–228 (2001).
- 28. Hersteinsson, P. & MacDonald, D. W. Interspecific Competition and the Geographical Distribution of Red and Arctic Foxes *Vulpes Vulpes* and *Alopex lagopus*. *Oikos* **64**, 505 (1992).
- 29. Service, C. N. *et al.* Indigenous Knowledge and Science Unite to Reveal Spatial and Temporal Dimensions of Distributional Shift in Wildlife of Conservation Concern. *PLoS ONE*

9, e101595 (2014).

- 30. Barry, J. P., Baxter, C. H., Sagarin, R. D. & Gilman, S. E. Climate-Related, Long-Term Faunal Changes in a California Rocky Intertidal Community. *Science* **267**, 672–675 (1995).
- 31. Grabherr, G., Gottfried, M. & Pauli, H. Climate effects on mountain plants. *Nature* **369**, 448–448 (1994).
- 32. Fredston-Hermann, A., Selden, R., Pinsky, M., Gaines, S. D. & Halpern, B. S. Cold range edges of marine fishes track climate change better than warm edges. *Glob. Change Biol.* **26**, 2908–2922 (2020).
- 33. Nye, J., Link, J., Hare, J. & Overholtz, W. Changing spatial distribution of fish stocks in relation to climate and population size on the Northeast United States continental shelf. *Mar. Ecol. Prog. Ser.* **393**, 111–129 (2009).
- 34. La Sorte, F. A. & Thompson III, F. R. POLEWARD SHIFTS IN WINTER RANGES OF NORTH AMERICAN BIRDS. *Ecology* **88**, 1803–1812 (2007).
- 35. Diamond, S. E., Frame, A. M., Martin, R. A. & Buckley, L. B. Species' traits predict phenological responses to climate change in butterflies. *Ecology* **92**, 1005–1012 (2011).
- 36. Shaffer, H. B., Fisher, R. N. & Davidson, C. The role of natural history collections in documenting species declines. *Trends Ecol. Evol.* **13**, 27–30 (1998).
- 37. Tingley, M. W. & Beissinger, S. R. Detecting range shifts from historical species occurrences: new perspectives on old data. *Trends Ecol. Evol.* **24**, 625–633 (2009).
- 38. Freeman, B. G., Scholer, M. N., Ruiz-Gutierrez, V. & Fitzpatrick, J. W. Climate change causes upslope shifts and mountaintop extirpations in a tropical bird community. *Proc. Natl. Acad. Sci.* **115**, 11982–11987 (2018).
- 39. Poloczanska, E. S. *et al.* Little change in the distribution of rocky shore faunal communities on the Australian east coast after 50 years of rapid warming. *J. Exp. Mar. Biol. Ecol.* **400**, 145–154 (2011).
- 40. Devictor, V. *et al.* Differences in the climatic debts of birds and butterflies at a continental scale. *Nat. Clim. Change* **2**, 121–124 (2012).
- 41. Lenoir, J., Gégout, J.-C., Pierrat, J.-C., Bontemps, J.-D. & Dhôte, J.-F. Differences between tree species seedling and adult altitudinal distribution in mountain forests during the recent warm period (1986–2006). *Ecography* **32**, 765–777 (2009).
- 42. Zhu, K., Woodall, C. W. & Clark, J. S. Failure to migrate: lack of tree range expansion in response to climate change. *Glob. Change Biol.* **18**, 1042–1052 (2012).
- 43. Monleon, V. J. & Lintz, H. E. Evidence of tree species' range shifts in a complex landscape. *PloS One* **10**, e0118069 (2015).
- 44. Hammerschlag, N. *et al.* Ocean warming alters the distributional range, migratory timing, and spatial protections of an apex predator, the tiger shark (Galeocerdo cuvier). *Glob. Change Biol.* **28**, 1990–2005 (2022).
- 45. Robinson, L. M. *et al.* Rapid assessment of an ocean warming hotspot reveals "high" confidence in potential species' range extensions. *Glob. Environ. Change* **31**, 28–37 (2015).
- 46. García Molinos, J. *et al.* Climate, currents and species traits contribute to early stages of marine species redistribution. *Commun. Biol.* **5**, 1–10 (2022).
- 47. Root, T. L. *et al.* Fingerprints of global warming on wild animals and plants. *Nature* **421**, 57–60 (2003).
- 48. Brooks, T. M. *et al.* Measuring Terrestrial Area of Habitat (AOH) and Its Utility for the IUCN

Red List. *Trends Ecol. Evol.* **34**, 977–986 (2019).

- 49. Fortin, M.-J. *et al.* Species' geographic ranges and distributional limits: pattern analysis and statistical issues. *Oikos* **108**, 7–17 (2005).
- 50. Gaston, K. J. & Fuller, R. A. The sizes of species' geographic ranges. *J. Appl. Ecol.* **46**, 1–9 (2009).
- 51. Gaston, K. J. *The Structure and Dynamics of Geographic Ranges*. (Oxford University Press, 2003).
- 52. Yalcin, S. & Leroux, S. J. Diversity and suitability of existing methods and metrics for quantifying species range shifts: Yalcin and Leroux. *Glob. Ecol. Biogeogr.* **26**, 609–624 (2017).
- 53. Shoo, L. P., Williams, S. E. & Hero, J.-M. Detecting climate change induced range shifts: Where and how should we be looking? *Austral Ecol.* **31**, 22–29 (2006).
- 54. Rubenstein, M. A. *et al.* Climate change and the global redistribution of biodiversity: substantial variation in empirical support for expected range shifts. *Environ. Evid.* **12**, 7 (2023).
- 55. Amano, T. *et al.* Tapping into non-English-language science for the conservation of global biodiversity. *PLOS Biol.* **19**, e3001296 (2021).
- 56. Konno, K. *et al.* Ignoring non-English-language studies may bias ecological meta-analyses. *Ecol. Evol.* **10**, 6373–6384 (2020).
- 57. Gervais, C. R., Champion, C. & Pecl, G. T. Species on the move around the Australian coastline: A continental-scale review of climate-driven species redistribution in marine systems. *Glob. Change Biol.* **27**, 3200–3217 (2021).
- 58. Thomas, C. D., Franco, A. M. A. & Hill, J. K. Range retractions and extinction in the face of climate warming. *Trends Ecol. Evol.* **21**, 415–416 (2006).
- 59. Freeman, B. G., Lee-Yaw, J. A., Sunday, J. M. & Hargreaves, A. L. Expanding, shifting and shrinking: The impact of global warming on species' elevational distributions. *Glob. Ecol. Biogeogr.* **27**, 1268–1276 (2018).
- 60. Doak, D. F. & Morris, W. F. Demographic compensation and tipping points in climateinduced range shifts. *Nature* **467**, 959–962 (2010).
- 61. Boisvert-Marsh, L., Périé, C. & de Blois, S. Shifting with climate? Evidence for recent changes in tree species distribution at high latitudes. *Ecosphere* **5**, art83 (2014).
- 62. Comte, L., Buisson, L., Daufresne, M. & Grenouillet, G. Climate-induced changes in the distribution of freshwater fish: observed and predicted trends: *Climate change and freshwater fish*. *Freshw. Biol.* **58**, 625–639 (2013).
- 63. Sunday, J. M., Bates, A. E. & Dulvy, N. K. Thermal tolerance and the global redistribution of animals. *Nat. Clim. Change* **2**, 686–690 (2012).
- 64. Pinsky, M. L., Eikeset, A. M., McCauley, D. J., Payne, J. L. & Sunday, J. M. Greater vulnerability to warming of marine versus terrestrial ectotherms. *Nature* **569**, 108–111 (2019).
- 65. Myers, A. A. Biogeographic Barriers and the Development of Marine Biodiversity. *Estuar. Coast. Shelf Sci.* **44**, 241–248 (1997).
- 66. Huang, Q. *et al.* Modeled distribution shifts of North American birds over four decades based on suitable climate alone do not predict observed shifts. *Sci. Total Environ.* **857**, 159603 (2023).
- 67. Lembrechts, J. J. & Lenoir, J. Microclimatic conditions anywhere at any time! *Glob. Change Biol.* **26**, 337–339 (2020).
- 68. Lenoir, J., Hattab, T. & Pierre, G. Climatic microrefugia under anthropogenic climate change: implications for species redistribution. *Ecography* **40**, 253–266 (2017).
- 69. Maclean, I. M. D. Predicting future climate at high spatial and temporal resolution. *Glob. Change Biol.* **26**, 1003–1011 (2020).
- 70. Zellweger, F., De Frenne, P., Lenoir, J., Rocchini, D. & Coomes, D. Advances in Microclimate Ecology Arising from Remote Sensing. *Trends Ecol. Evol.* **34**, 327–341 (2019).
- 71. Haesen, S. *et al.* ForestTemp Sub-canopy microclimate temperatures of European forests. *Glob. Change Biol.* **27**, 6307–6319 (2021).
- 72. Lembrechts, J. J. *et al.* Global maps of soil temperature. *Glob. Change Biol.* **28**, 3110–3144 (2022).
- 73. Maclean, I. M. D. & Early, R. Macroclimate data overestimate range shifts of plants in response to climate change. *Nat. Clim. Change* (2023) doi:10.1038/s41558-023-01650-3.
- 74. Dobrowski, S. Z. *et al.* The climate velocity of the contiguous United States during the 20th century. *Glob. Change Biol.* **19**, 241–251 (2013).
- 75. Crimmins, S. M., Dobrowski, S. Z., Greenberg, J. A., Abatzoglou, J. T. & Mynsberge, A. R. Changes in Climatic Water Balance Drive Downhill Shifts in Plant Species' Optimum Elevations. *Science* **331**, 324–327 (2011).
- 76. Hordley, L. A., Fox, R., Suggitt, A. J. & Bourn, N. A. D. Precipitation buffers temperaturedriven local extinctions of moths at warm range margins. *Ecol. Lett.* **00**, 1–11 (2023).
- 77. Tingley, M. W., Monahan, W. B., Beissinger, S. R. & Moritz, C. Birds track their Grinnellian

niche through a century of climate change. *Proc. Natl. Acad. Sci.* **106**, 19637–19643 (2009).

- 78. McHenry, J., Welch, H., Lester, S. E. & Saba, V. Projecting marine species range shifts from only temperature can mask climate vulnerability. *Glob. Change Biol.* **25**, 4208–4221 (2019).
- 79. Spence, A. R. & Tingley, M. W. The challenge of novel abiotic conditions for species undergoing climate-induced range shifts. *Ecography* **43**, 1571–1590 (2020).
- 80. Huffeldt, N. P. Photic Barriers to Poleward Range-shifts. *Trends Ecol. Evol.* **35**, 652–655 (2020).
- 81. Muir, P. R., Wallace, C. C., Done, T. & Aguirre, J. D. Limited scope for latitudinal extension of reef corals. *Science* **348**, 1135–1138 (2015).
- 82. Jacobsen, D. The dilemma of altitudinal shifts: caught between high temperature and low oxygen. *Front. Ecol. Environ.* **18**, 211–218 (2020).
- 83. Sunday, J. M. *et al.* Biological sensitivities to high-resolution climate change projections in the California current marine ecosystem. *Glob. Change Biol.* **28**, 5726–5740 (2022).
- 84. Twiname, S. *et al.* Mismatch of thermal optima between performance measures, life stages and species of spiny lobster. *Sci. Rep.* **10**, 21235 (2020).
- 85. Harris, R. M. B. *et al.* Biological responses to the press and pulse of climate trends and extreme events. *Nat. Clim. Change* **8**, 579–587 (2018).
- 86. Smith, M. D. An ecological perspective on extreme climatic events: a synthetic definition and framework to guide future research. *J. Ecol.* **99**, 656–663 (2011).
- 87. Zong, S. *et al.* Upward range shift of a dominant alpine shrub related to 50 years of snow cover change. *Remote Sens. Environ.* **268**, 112773 (2022).
- 88. Edgar, G. J. *et al.* Continent-wide declines in shallow reef life over a decade of ocean warming. *Nature* **615**, 858–865 (2023).
- 89. Dullinger, S. *et al.* Post-glacial migration lag restricts range filling of plants in the European Alps. *Glob. Ecol. Biogeogr.* **21**, 829–840 (2012).
- 90. Carvajal-Quintero, J. *et al.* Drainage network position and historical connectivity explain global patterns in freshwater fishes' range size. *Proc. Natl. Acad. Sci.* **116**, 13434–13439 (2019).
- 91. Matthews, W. J. & Zimmerman, E. G. Potential Effects of Global Warming on Native Fishes of the Southern Great Plains and the Southwest. *Fisheries* **15**, 26–32 (1990).
- 92. Guo, F., Lenoir, J. & Bonebrake, T. C. Land-use change interacts with climate to determine elevational species redistribution. *Nat. Commun.* **9**, 1315 (2018).
- 93. Hill, J. K., Thomas, C. D. & Huntley, B. Climate and habitat availability determine 20th century changes in a butterfly's range margin. *Proc. R. Soc. Lond. B Biol. Sci.* **266**, 1197– 1206 (1999).
- 94. Morelli, T. L. *et al.* Anthropogenic refugia ameliorate the severe climate-related decline of a montane mammal along its trailing edge. *Proc. R. Soc. B Biol. Sci.* **279**, 4279–4286 (2012).
- 95. Fischer, J. & Lindenmayer, D. B. Landscape modification and habitat fragmentation: a synthesis. *Glob. Ecol. Biogeogr.* **16**, 265–280 (2007).
- 96. Schloss, C. A., Nuñez, T. A. & Lawler, J. J. Dispersal will limit ability of mammals to track climate change in the Western Hemisphere. *Proc. Natl. Acad. Sci.* **109**, 8606–8611 (2012).
- 97. Hodgson, J. A., Randle, Z., Shortall, C. R. & Oliver, T. H. Where and why are species' range shifts hampered by unsuitable landscapes? *Glob. Change Biol.* **28**, 4765–4774 (2022).
- 98. Thomas, C. D. *et al.* Protected areas facilitate species' range expansions. *Proc. Natl. Acad. Sci.* **109**, 14063–14068 (2012).
- 99. Coolen, J. W. P. *et al.* Marine stepping-stones: Connectivity of Mytilus edulis populations between offshore energy installations. *Mol. Ecol.* **29**, 686–703 (2020).
- 100. Reino, L. *et al.* Modelling landscape constraints on farmland bird species range shifts under climate change. *Sci. Total Environ.* **625**, 1596–1605 (2018).
- 101. Lawler, J. J., Ruesch, A. S., Olden, J. D. & McRae, B. H. Projected climate-driven faunal movement routes. *Ecol. Lett.* **16**, 1014–1022 (2013).
- 102. Littlefield, C. E., Krosby, M., Michalak, J. L. & Lawler, J. J. Connectivity for species on the move: supporting climate-driven range shifts. *Front. Ecol. Environ.* **17**, 270–278 (2019).
- 103. García Molinos, J., Burrows, M. T. & Poloczanska, E. S. Ocean currents modify the coupling between climate change and biogeographical shifts. *Sci. Rep.* **7**, 1332 (2017).
- 104. Kling, M. M. & Ackerly, D. D. Global wind patterns and the vulnerability of wind-dispersed species to climate change. *Nat. Clim. Change* **10**, 868–875 (2020).
- 105. Copeland, S. M., Bradford, J. B., Duniway, M. C. & Butterfield, B. J. Life history characteristics may be as important as climate projections for defining range shifts: An example for common tree species in the intermountain western US. *Divers. Distrib.* **24**, 1844–1859 (2018).
- 106. Naoe, S. *et al.* Downhill seed dispersal by temperate mammals: a potential threat to plant escape from global warming. *Sci. Rep.* **9**, 14932 (2019).
- 107. Ramos, J. E. *et al.* Population genetic signatures of a climate change driven marine range extension. *Sci. Rep.* **8**, 9558 (2018).
- 108. MacLean, S. A. & Beissinger, S. R. Species' traits as predictors of range shifts under contemporary climate change: A review and meta-analysis. *Glob. Change Biol.* **23**, 4094– 4105 (2017).
- 109. Sunday, J. M. *et al.* Species traits and climate velocity explain geographic range shifts in an ocean-warming hotspot. *Ecol. Lett.* **18**, 944–953 (2015).
- 110. Angert, A. L. *et al.* Do species' traits predict recent shifts at expanding range edges?: Traits and range shifts. *Ecol. Lett.* **14**, 677–689 (2011).
- 111. Alofs, K. M., Jackson, D. A. & Lester, N. P. Ontario freshwater fishes demonstrate differing range-boundary shifts in a warming climate. *Divers. Distrib.* **20**, 123–136 (2014).
- 112. Beissinger, S. R. & Riddell, E. A. Why Are Species' Traits Weak Predictors of Range Shifts? *Annu. Rev. Ecol. Evol. Syst.* **52**, 47–66 (2021).
- 113. HilleRisLambers, J., Harsch, M. A., Ettinger, A. K., Ford, K. R. & Theobald, E. J. How will biotic interactions influence climate change–induced range shifts? *Ann. N. Y. Acad. Sci.* **1297**, 112–125 (2013).
- 114. Kerner, J. M., Krauss, J., Maihoff, F., Bofinger, L. & Classen, A. Alpine butterflies want to fly high: Species and communities shift upwards faster than their host plants. *Ecology* **104**, e3848 (2023).
- 115. Merrill, R. M. *et al.* Combined effects of climate and biotic interactions on the elevational range of a phytophagous insect. *J. Anim. Ecol.* **77**, 145–155 (2008).
- 116. Afkhami, M. E., McIntyre, P. J. & Strauss, S. Y. Mutualist-mediated effects on species' range limits across large geographic scales. *Ecol. Lett.* **17**, 1265–1273 (2014).
- 117. Lankau, R. A., Zhu, K. & Ordonez, A. Mycorrhizal strategies of tree species correlate with

trailing range edge responses to current and past climate change. *Ecology* **96**, 1451–1458 (2015).

- 118. Sirén, A. P. K. & Morelli, T. L. Interactive range-limit theory (iRLT): An extension for predicting range shifts. *J. Anim. Ecol.* **89**, 940–954 (2020).
- 119. Pokallus, J. W. & Pauli, J. . N. Population dynamics of a northern-adapted mammal: disentangling the influence of predation and climate change. *Ecol. Appl.* **25**, 1546–1556 (2015).
- 120. Wallingford, P. D. *et al.* Adjusting the lens of invasion biology to focus on the impacts of climate-driven range shifts. *Nat. Clim. Change* **10**, 398–405 (2020).
- 121. Harley, C. D. G. Climate Change, Keystone Predation, and Biodiversity Loss. *Science* **334**, 1124–1127 (2011).
- 122. Lenoir, J. *et al.* Going against the flow: potential mechanisms for unexpected downslope range shifts in a warming climate. *Ecography* **33**, 295–303 (2010).
- 123. Pollock, L. J. *et al.* Understanding co-occurrence by modelling species simultaneously with a Joint Species Distribution Model (JSDM). *Methods Ecol. Evol.* **5**, 397–406 (2014).
- 124. Devarajan, K., Morelli, T. L. & Tenan, S. Multi-species occupancy models: review, roadmap, and recommendations. *Ecography* **43**, 1612–1624 (2020).
- 125. Zhang, C., Chen, Y., Xu, B., Xue, Y. & Ren, Y. Improving prediction of rare species' distribution from community data. *Sci. Rep.* **10**, 12230 (2020).
- 126. Wisz, M. S. *et al.* The role of biotic interactions in shaping distributions and realised assemblages of species: implications for species distribution modelling. *Biol. Rev.* **88**, 15– 30 (2013).
- 127. Diamond, S. E. Contemporary climate-driven range shifts: Putting evolution back on the table. *Funct. Ecol.* **32**, 1652–1665 (2018).
- 128. Dudaniec, R. Y. *et al.* Latitudinal clines in sexual selection, sexual size dimorphism and sexspecific genetic dispersal during a poleward range expansion. *J. Anim. Ecol.* **91**, 1104–1118 (2022).
- 129. Krause, J. S. *et al.* Breeding on the leading edge of a northward range expansion: differences in morphology and the stress response in the arctic Gambel's white-crowned sparrow. *Oecologia* **180**, 33–44 (2016).
- 130. Campbell-Staton, S. C. *et al.* Parallel selection on thermal physiology facilitates repeated adaptation of city lizards to urban heat islands. *Nat. Ecol. Evol.* **4**, 652–658 (2020).
- 131. Diamond, S. E., Chick, L., Perez, A., Strickler, S. A. & Martin, R. A. Rapid evolution of ant thermal tolerance across an urban-rural temperature cline. *Biol. J. Linn. Soc.* **121**, 248–257 (2017).
- 132. Donelson, J. M. *et al.* Understanding interactions between plasticity, adaptation and range shifts in response to marine environmental change. *Philos. Trans. R. Soc. B Biol. Sci.* **374**, 20180186 (2019).
- 133. Bridle, J. R. & Vines, T. H. Limits to evolution at range margins: when and why does adaptation fail? *Trends Ecol. Evol.* **22**, 140–147 (2007).
- 134. Nadeau, C. P. & Urban, M. C. Eco-evolution on the edge during climate change. *Ecography* **42**, 1280–1297 (2019).
- 135. Guo, Q., Taper, M., Schoenberger, M. & Brandle, J. Spatial-Temporal Population Dynamics across Species Range: From Centre to Margin. *Oikos* **108**, 47–57 (2005).
- 136. Taheri, S., Naimi, B., Rahbek, C. & Araújo, M. B. Improvements in reports of species redistribution under climate change are required. *Sci. Adv.* **7**, eabe1110 (2021).
- 137. Iseli, E. *et al.* Rapid upwards spread of non-native plants in mountains across continents. *Nat. Ecol. Evol.* **7**, 405–413 (2023).
- 138. Mazalla, L. & Diekmann, M. Regression to the mean in vegetation science. *J. Veg. Sci.* **33**, e13117 (2022).
- 139. Bates, A. E. *et al.* Distinguishing geographical range shifts from artefacts of detectability and sampling effort. *Divers. Distrib.* **21**, 13–22 (2015).
- 140. Davis, A. *et al. Tools for predicting wildlife species distribution response to ecological shifts*. https://hdl.handle.net/11681/33482 (2019) doi:10.21079/11681/33482.
- 141. Lindsay, K. *Wildlife conservation, protected areas and climate change in Canada: implications of projected species range shifts*. (Canadian Council on Ecological Areas, 2016).
- 142. Weiskopf, S. R. *et al.* Climate change effects on biodiversity, ecosystems, ecosystem services, and natural resource management in the United States. *Sci. Total Environ.* **733**, 137782 (2020).
- 143. Pinsky, M. L. *et al.* Preparing ocean governance for species on the move. *Science* **360**, 1189–1191 (2018).
- 144. Scheffers, B. R. & Pecl, G. Persecuting, protecting or ignoring biodiversity under climate change. *Nat. Clim. Change* **9**, 581–586 (2019).
- 145. Bowler, D. E. *et al.* Cross-realm assessment of climate change impacts on species' abundance trends. *Nat. Ecol. Evol.* **1**, 1–7 (2017).
- 146. Feeley, K. J., Bravo-Avila, C., Fadrique, B., Perez, T. M. & Zuleta, D. Climate-driven changes in the composition of New World plant communities. *Nat. Clim. Change* **10**, 965–970 (2020).
- 147. Hughes, T. P. *et al.* Global warming transforms coral reef assemblages. *Nature* **556**, 492– 496 (2018).
- 148. Kuhn, E. & Gégout, J.-C. Highlighting declines of cold-demanding plant species in lowlands under climate warming. *Ecography* **42**, 36–44 (2019).
- 149. Kumagai, N. H. *et al.* Ocean currents and herbivory drive macroalgae-to-coral community shift under climate warming. *Proc. Natl. Acad. Sci.* **115**, 8990–8995 (2018).
- 150. Balik, J. A., Greig, H. S., Taylor, B. W. & Wissinger, S. A. Consequences of climate-induced range expansions on multiple ecosystem functions. *Commun. Biol.* **6**, 390 (2023).
- 151. Cauvy-Fraunié, S. & Dangles, O. A global synthesis of biodiversity responses to glacier retreat. *Nat. Ecol. Evol.* **3**, 1675–1685 (2019).
- 152. Stuart-Smith, R. D., Mellin, C., Bates, A. E. & Edgar, G. J. Habitat loss and range shifts contribute to ecological generalization among reef fishes. *Nat. Ecol. Evol.* **5**, 656–662 (2021).
- 153. Bertrand, R. *et al.* Changes in plant community composition lag behind climate warming in lowland forests. *Nature* **479**, 517–520 (2011).
- 154. Davis, T. R., Knott, N. A., Champion, C. & Przeslawski, R. Impacts of Climate Change on Densities of the Urchin Centrostephanus rodgersii Vary among Marine Regions in Eastern Australia. *Diversity* **15**, 419 (2023).
- 155. Ling, S. D. Range expansion of a habitat-modifying species leads to loss of taxonomic

diversity: a new and impoverished reef state. *Oecologia* **156**, 883–894 (2008).

- 156. Ling, S. & Johnson, C. Population dynamics of an ecologically important range-extender: kelp beds versus sea urchin barrens. *Mar. Ecol. Prog. Ser.* **374**, 113–125 (2009).
- 157. Fazlioglu, F., Wan, J. S. H. & Chen, L. Latitudinal shifts in mangrove species worldwide: evidence from historical occurrence records. *Hydrobiologia* **847**, 4111–4123 (2020).
- 158. Nagelkerken, I. *et al.* The habitat function of mangroves for terrestrial and marine fauna: A review. *Aquat. Bot.* **89**, 155–185 (2008).
- 159. Descombes, P. *et al.* Novel trophic interactions under climate change promote alpine plant coexistence. *Science* **370**, 1469–1473 (2020).
- 160. Pecl, G. T. *et al.* Autonomous adaptation to climate-driven change in marine biodiversity in a global marine hotspot. *Ambio* **48**, 1498–1515 (2019).
- 161. Oremus, K. L. Climate variability reduces employment in New England fisheries. *Proc. Natl. Acad. Sci.* **116**, 26444–26449 (2019).
- 162. Young, T. *et al.* Adaptation strategies of coastal fishing communities as species shift poleward. *ICES J. Mar. Sci.* **76**, 93–103 (2019).
- 163. Sloat, L. L. *et al.* Climate adaptation by crop migration. *Nat. Commun.* **11**, 1243 (2020).
- 164. Bebber, D. P., Ramotowski, M. A. T. & Gurr, S. J. Crop pests and pathogens move polewards in a warming world. *Nat. Clim. Change* **3**, 985–988 (2013).
- 165. Blackfeet Nation. Blackfeet climate change adaptation plan. (2018).
- 166. Sweet, L. C. *et al.* Congruence between future distribution models and empirical data for an iconic species at Joshua Tree National Park. *Ecosphere* **10**, e02763 (2019).
- 167. Carlson, C. J., Bannon, E., Mendenhall, E., Newfield, T. & Bansal, S. Rapid range shifts in

African Anopheles mosquitoes over the last century. *Biol. Lett.* **19**, 20220365 (2023).

- 168. Clow, K. M. *et al.* Northward range expansion of Ixodes scapularis evident over a short timescale in Ontario, Canada. *PLOS ONE* **12**, e0189393 (2017).
- 169. McCracken, G. F. *et al.* Rapid range expansion of the Brazilian free-tailed bat in the southeastern United States, 2008–2016. *J. Mammal.* **99**, 312–320 (2018).
- 170. Siraj, A. S. *et al.* Altitudinal changes in malaria incidence in highlands of Ethiopia and Colombia. *Science* **343**, 1154–1158 (2014).
- 171. Alkishe, A., Raghavan, R. K. & Peterson, A. T. Likely Geographic Distributional Shifts among Medically Important Tick Species and Tick-Associated Diseases under Climate Change in North America: A Review. *Insects* **12**, 225 (2021).
- 172. Ryan, S. J. *et al.* Warming temperatures could expose more than 1.3 billion new people to Zika virus risk by 2050. *Glob. Change Biol.* **27**, 84–93 (2021).
- 173. Chinain, M., Gatti, C. M. i., Darius, H. T., Quod, J.-P. & Tester, P. A. Ciguatera poisonings: A global review of occurrences and trends. *Harmful Algae* **102**, 101873 (2021).
- 174. Kulkarni, M. A. *et al.* 10 Years of Environmental Change on the Slopes of Mount Kilimanjaro and Its Associated Shift in Malaria Vector Distributions. *Front. Public Health* **4**, (2016).
- 175. Ryan, S. J., Carlson, C. J., Mordecai, E. A. & Johnson, L. R. Global expansion and redistribution of Aedes-borne virus transmission risk with climate change. *PLoS Negl. Trop. Dis.* **13**, e0007213 (2019).
- 176. Carlson, C. J. *et al.* Climate change increases cross-species viral transmission risk. *Nature* **607**, 555–562 (2022).
- 177. Gray, C. L. *et al.* Local biodiversity is higher inside than outside terrestrial protected areas worldwide. *Nat. Commun.* **7**, 12306 (2016).
- 178. Bruno, J. F. *et al.* Climate change threatens the world's marine protected areas. *Nat. Clim. Change* **8**, 499–503 (2018).
- 179. Lewis, S. A., Stortini, C. H., Boyce, D. G. & Stanley, R. R. E. Climate change, species thermal emergence, and conservation design: a case study in the Canadian Northwest Atlantic. *FACETS* **8**, 1–16 (2023).
- 180. Parks, S. A., Holsinger, L. M., Abatzoglou, J. T., Littlefield, C. E. & Zeller, K. A. Protected areas not likely to serve as steppingstones for species undergoing climate-induced range shifts. *Glob. Change Biol.* **00**, 1–16 (2023).
- 181. D'Aloia, C. C. *et al.* Coupled Networks of Permanent Protected Areas and Dynamic Conservation Areas for Biodiversity Conservation Under Climate Change. *Front. Ecol. Evol.* **7**, (2019).
- 182. Cashion, T. *et al.* Shifting seas, shifting boundaries: Dynamic marine protected area designs for a changing climate. *PLOS ONE* **15**, e0241771 (2020).
- 183. Pinsky, M. L., Rogers, L. A., Morley, J. W. & Frölicher, T. L. Ocean planning for species on the move provides substantial benefits and requires few trade-offs. *Sci. Adv.* **6**, eabb8428 (2020).
- 184. Lewison, R. *et al.* Dynamic Ocean Management: Identifying the Critical Ingredients of Dynamic Approaches to Ocean Resource Management. *BioScience* **65**, 486–498 (2015).
- 185. Maxwell, S. M. *et al.* Dynamic ocean management: Defining and conceptualizing real-time management of the ocean. *Mar. Policy* **58**, 42–50 (2015).
- 186. Golet, G. H. *et al.* Using ricelands to provide temporary shorebird habitat during migration. *Ecol. Appl.* **28**, 409–426 (2018).
- 187. Reynolds, M. D. *et al.* Dynamic conservation for migratory species. *Sci. Adv.* **3**, e1700707 (2017).
- 188. Bloom, A. The Cumberland Forest Project: 253,000 Acres of Preserved Land. https://www.nature.org/en-us/what-we-do/our-priorities/protect-water-and-land/landand-water-stories/cumberland-forest-project (2019).
- 189. Bentrup, G. *Conservation Buffers—Design guidelines for buffers, corridors, and greenways*. SRS-GTR-109 https://www.fs.usda.gov/treesearch/pubs/33522 (2008) doi:10.2737/SRS-GTR-109.
- 190. Handler, S., Pike, C., St. Clair, B., Abbotts, H. & Janowiak, M. Assisted Migration. USDA Forest Service Climate Change Resource Center.

https://www.fs.usda.gov/ccrc/topics/assisted-migration (2018).

- 191. Ricciardi, A. & Simberloff, D. Assisted colonization is not a viable conservation strategy. *Trends Ecol. Evol.* **24**, 248–253 (2009).
- 192. Carbajal-Navarro, A. *et al.* Ecological Restoration of Abies religiosa Forests Using Nurse Plants and Assisted Migration in the Monarch Butterfly Biosphere Reserve, Mexico. *Front. Ecol. Evol.* **7**, 421 (2019).
- 193. Widhalm, T., Fourcade, Y., Frank, T. & Öckinger, E. Population dynamics of the butterfly Pyrgus armoricanus after translocation beyond its northern range margin. *Insect Conserv. Divers.* **13**, 617–629 (2020).
- 194. Liu, H. *et al.* Overcoming extreme weather challenges: Successful but variable assisted

colonization of wild orchids in southwestern China. *Biol. Conserv.* **150**, 68–75 (2012).

- 195. Twardek, W. M. *et al.* The application of assisted migration as a climate change adaptation tactic: An evidence map and synthesis. *Biol. Conserv.* **280**, 109932 (2023).
- 196. Walsworth, T. E. *et al.* Management for network diversity speeds evolutionary adaptation to climate change. *Nat. Clim. Change* **9**, 632–636 (2019).
- 197. Norberg, J., Urban, M. C., Vellend, M., Klausmeier, C. A. & Loeuille, N. Eco-evolutionary responses of biodiversity to climate change. *Nat. Clim. Change* **2**, 747–751 (2012).
- 198. Palacios-Abrantes, J. *et al.* Quantifying fish range shifts across poorly defined management boundaries. *PLOS ONE* **18**, e0279025 (2023).
- 199. Spijkers, J. & Boonstra, W. J. Environmental change and social conflict: the northeast Atlantic mackerel dispute. *Reg. Environ. Change* **17**, 1835–1851 (2017).
- 200. Parks, S. A. *et al.* Efficacy of the global protected area network is threatened by disappearing climates and potential transboundary range shifts. *Environ. Res. Lett.* **17**, 054016 (2022).
- 201. Palacios-Abrantes, J. *et al.* Timing and magnitude of climate-driven range shifts in transboundary fish stocks challenge their management. *Glob. Change Biol.* **28**, 2312–2326 (2022).
- 202. Titley, M. A., Butchart, S. H. M., Jones, V. R., Whittingham, M. J. & Willis, S. G. Global inequities and political borders challenge nature conservation under climate change. *Proc. Natl. Acad. Sci.* **118**, e2011204118 (2021).
- 203. Velazco, S. J. E., Villalobos, F., Galvão, F. & De Marco Júnior, P. A dark scenario for Cerrado plant species: Effects of future climate, land use and protected areas ineffectiveness.

Divers. Distrib. **25**, 660–673 (2019).

- 204. Xu, C., Kohler, T. A., Lenton, T. M., Svenning, J.-C. & Scheffer, M. Future of the human climate niche. *Proc. Natl. Acad. Sci.* **117**, 11350–11355 (2020).
- 205. Lenton, T. M. *et al.* Quantifying the human cost of global warming. *Nat. Sustain.* (2023) doi:10.1038/s41893-023-01132-6.
- 206. Moat, J. *et al.* Resilience potential of the Ethiopian coffee sector under climate change. *Nat. Plants* **3**, 1–14 (2017).
- 207. Pham, Y., Reardon-Smith, K., Mushtaq, S. & Cockfield, G. The impact of climate change and variability on coffee production: a systematic review. *Clim. Change* **156**, 609–630 (2019).
- 208. Swartz, W., Sala, E., Tracey, S., Watson, R. & Pauly, D. The Spatial Expansion and Ecological Footprint of Fisheries (1950 to Present). *PLOS ONE* **5**, e15143 (2010).
- 209. Cao, Y. *et al.* Trans-Arctic shipping routes expanding faster than the model projections. *Glob. Environ. Change* **73**, 102488 (2022).
- 210. Paglia, E. A higher level of civilisation? The transformation of Ny-Ålesund from Arctic coalmining settlement in Svalbard to global environmental knowledge center at 79° North. *Polar Rec.* **56**, e15 (2020).
- 211. Sokolickova, Z., Meyer, A. & Vlakhov, A. V. Changing Svalbard: Tracing interrelated socioeconomic and environmental change in remote Arctic settlements. *Polar Rec.* **58**, e23 (2022).
- 212. OECD. *Climate Change in the European Alps: Adapting Winter Tourism and Natural Hazards Management*. (Organisation for Economic Co-operation and Development, 2007).
- 213. Ma, S., Craig, C. A. & Feng, S. The Camping Climate Index (CCI): The development,

validation, and application of a camping-sector tourism climate index. *Tour. Manag.* **80**, 104105 (2020).

- 214. Fisichelli, N. A., Schuurman, G. W., Monahan, W. B. & Ziesler, P. S. Protected Area Tourism in a Changing Climate: Will Visitation at US National Parks Warm Up or Overheat? *PLOS ONE* **10**, e0128226 (2015).
- 215. Grillakis, M. G., Koutroulis, A. G., Seiradakis, K. D. & Tsanis, I. K. Implications of 2 °C global warming in European summer tourism. *Clim. Serv.* **1**, 30–38 (2016).
- 216. World Tourism Organization, U. N. E. P. Climate Change and Tourism: Responding to Global Challenges. (2008).
- 217. Zhang, Y., Held, I. & Fueglistaler, S. Projections of tropical heat stress constrained by atmospheric dynamics. *Nat. Geosci.* **14**, 133–137 (2021).
- 218. Lindenmayer, D. B., Lavery, T. & Scheele, B. C. Why We Need to Invest in Large-Scale, Long-Term Monitoring Programs in Landscape Ecology and Conservation Biology. *Curr. Landsc. Ecol. Rep.* **7**, 137–146 (2022).
- 219. Maclean, I. M. D. & Wilson, R. J. Recent ecological responses to climate change support predictions of high extinction risk. *Proc. Natl. Acad. Sci.* **108**, 12337–12342 (2011).
- 220. Kissling, W. D. *et al.* Building essential biodiversity variables (EBV s) of species distribution and abundance at a global scale. *Biol. Rev.* **93**, 600–625 (2018).
- 221. Gonzalez, A. *et al.* A global biodiversity observing system to unite monitoring and guide action. *Nat. Ecol. Evol.* (2023) doi:10.1038/s41559-023-02171-0.
- 222. Chandler, M. *et al.* Contribution of citizen science towards international biodiversity monitoring. *Biol. Conserv.* **213**, 280–294 (2017).
- 223. Pecl, G. T. *et al.* Redmap Australia: Challenges and Successes With a Large-Scale Citizen Science-Based Approach to Ecological Monitoring and Community Engagement on Climate Change. *Front. Mar. Sci.* **6**, 349 (2019).
- 224. Middleton, I. *et al.* Introduced alien, range extension or just visiting? Combining citizen science observations and expert knowledge to classify range dynamics of marine fishes. *Divers. Distrib.* **27**, 1278–1293 (2021).
- 225. Fraisl, D. *et al.* Citizen science in environmental and ecological sciences. *Nat. Rev. Methods Primer* **2**, 64 (2022).
- 226. Paprocki, N., Heath, J. A. & Novak, S. J. Regional Distribution Shifts Help Explain Local Changes in Wintering Raptor Abundance: Implications for Interpreting Population Trends. *PLOS ONE* **9**, 9 (2014).
- 227. Wilson, S., Anderson, E. M., Wilson, A. S. G., Bertram, D. F. & Arcese, P. Citizen Science Reveals an Extensive Shift in the Winter Distribution of Migratory Western Grebes. *PLoS ONE* **8**, e65408 (2013).
- 228. Kirchman, J. J. & Van Keuren, A. E. Altitudinal Range Shifts of Birds At the Southern Periphery of the Boreal Forest: 40 Years of Change In the Adirondack Mountains. *Wilson J. Ornithol.* **129**, 742–753 (2017).
- 229. Bird, T. J. *et al.* Statistical solutions for error and bias in global citizen science datasets. *Biol. Conserv.* **173**, 144–154 (2014).
- 230. Estrada, A., Morales-Castilla, I., Caplat, P. & Early, R. Usefulness of Species Traits in Predicting Range Shifts. *Trends Ecol. Evol.* **31**, 190–203 (2016).
- 231. Green, S. J., Brookson, C. B., Hardy, N. A. & Crowder, L. B. Trait-based approaches to global

change ecology: moving from description to prediction. *Proc. R. Soc. B Biol. Sci.* **289**, 20220071 (2022).

- 232. Couet, J. *et al.* Short-lived species move uphill faster under climate change. *Oecologia* **198**, 877–888 (2022).
- 233. Tekwa, E. W., Watson, J. R. & Pinsky, M. L. Body size and food–web interactions mediate species range shifts under warming. *Proc. R. Soc. B Biol. Sci.* **289**, 20212755 (2022).
- 234. Brito-Morales, I. *et al.* Climate Velocity Can Inform Conservation in a Warming World. *Trends Ecol. Evol.* **33**, 441–457 (2018).
- 235. McRae, B. H., Viral B. Shah & Edelman, A. Circuitscape: modeling landscape connectivity to promote conservation and human health. (2016) doi:10.13140/RG.2.1.4265.1126.
- 236. Bocedi, G. *et al.* RangeShifter 2.0: an extended and enhanced platform for modelling spatial eco-evolutionary dynamics and species' responses to environmental changes. *Ecography* **44**, 1453–1462 (2021).
- 237. Cotto, O., Schmid, M. & Guillaume, F. Nemo-age: Spatially explicit simulations of ecoevolutionary dynamics in stage-structured populations under changing environments. *Methods Ecol. Evol.* **11**, 1227–1236 (2020).
- 238. Kearney, M. R. & Porter, W. P. NicheMapR an R package for biophysical modelling: the ectotherm and Dynamic Energy Budget models. *Ecography* **43**, 85–96 (2020).
- 239. Bush, A. *et al.* Incorporating evolutionary adaptation in species distribution modelling reduces projected vulnerability to climate change. *Ecol. Lett.* **19**, 1468–1478 (2016).
- 240. Engler, R., Hordijk, W. & Guisan, A. The MIGCLIM R package seamless integration of dispersal constraints into projections of species distribution models. *Ecography* **35**, 872–

878 (2012).

- 241. Evans, M. E. K., Merow, C., Record, S., McMahon, S. M. & Enquist, B. J. Towards Processbased Range Modeling of Many Species. *Trends Ecol. Evol.* **31**, 860–871 (2016).
- 242. Urban, M. C., Tewksbury, J. J. & Sheldon, K. S. On a collision course: competition and dispersal differences create no-analogue communities and cause extinctions during climate change. *Proc. R. Soc. B Biol. Sci.* **279**, 2072–2080 (2012).
- 243. Bleyhl, B. *et al.* Reducing persecution is more effective for restoring large carnivores than restoring their prey. *Ecol. Appl.* **31**, e02338 (2021).
- 244. Dietze, M. C. *et al.* Iterative near-term ecological forecasting: Needs, opportunities, and challenges. *Proc. Natl. Acad. Sci.* **115**, 1424–1432 (2018).
- 245. Halpern, B. S. *et al.* Priorities for synthesis research in ecology and environmental science. *Ecosphere* **14**, (2023).
- 246. Eveson, J. P., Hobday, A. J., Hartog, J. R., Spillman, C. M. & Rough, K. M. Seasonal forecasting of tuna habitat in the Great Australian Bight. *Fish. Res.* **170**, 39–49 (2015).
- 247. Howell, E. A. *et al.* Enhancing the TurtleWatch product for leatherback sea turtles, a dynamic habitat model for ecosystem-based management. *Fish. Oceanogr.* **24**, 57–68 (2015).
- 248. Hobday, A. J., Spillman, C. M., Paige Eveson, J. & Hartog, J. R. Seasonal forecasting for decision support in marine fisheries and aquaculture. *Fish. Oceanogr.* **25**, 45–56 (2016).
- 249. Hobday, A. J., Hartog, J. R., Spillman, C. M. & Alves, O. Seasonal forecasting of tuna habitat for dynamic spatial management. *Can. J. Fish. Aquat. Sci.* **68**, 898–911 (2011).
- 250. Horton, K. G., Van Doren, B. M., Albers, H. J., Farnsworth, A. & Sheldon, D. Near-term

ecological forecasting for dynamic aeroconservation of migratory birds. *Conserv. Biol.* **35**,

1777–1786 (2021).